



Pyrolysis of solid waste and its components in a lab scale induction-heating reactor

Oscar Sosa Sabogal, Sylvie Valin, Sébastien Thiery, Sylvain Salvador

► To cite this version:

Oscar Sosa Sabogal, Sylvie Valin, Sébastien Thiery, Sylvain Salvador. Pyrolysis of solid waste and its components in a lab scale induction-heating reactor. *Detritus*, 2021, 15, pp.107-112. 10.31025/2611-4135/2021.15094 . hal-03328491

HAL Id: hal-03328491

<https://imt-mines-albi.hal.science/hal-03328491>

Submitted on 30 Aug 2021

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.



Distributed under a Creative Commons Attribution - NonCommercial - NoDerivatives 4.0 International License

PYROLYSIS OF SOLID WASTE AND ITS COMPONENTS IN A LAB SCALE INDUCTION-HEATING REACTOR

Oscar Sosa Sabogal ^{1,*}, Sylvie Valin ¹, Sébastien Thiery ¹ and Sylvain Salvador ²

¹ University Grenoble Alpes, CEA, LITEN, DTBH, L2CS; 17 rue des Martyrs, Grenoble, 38000, France

² Centre RAPSODEE, IMT Mines Albi, CNRS UMR 5302, Allée des sciences, Albi, 81013, France

Article Info:

Received:
22 February 2021
Revised:
22 April 2021
Accepted:
17 May 2021
Available online:
30 June 2021

Keywords:

Pyrolysis
Gasification
Solid waste
Induction heating
Waste derived fuel

ABSTRACT

The present study investigates the thermochemical conversion of Solid Recovered Fuel (SRF), represented by selected "model materials". A laboratory-scale induction heated device was specifically developed to achieve fast pyrolysis conditions close to those encountered in a fluidized bed reactor. The novel device can handle up to 5 grams of solid, allowing fast heating rates (near 70°C/s) and a homogeneous distribution of temperature all along the reactor. Pyrolysis tests of a SRF sample and four model materials (Polyethylene, Polyethylene Terephthalate, beech wood, cardboard) were performed at 800°C. The yield and composition of the produced gas for each sample were determined. Experimental results will help to elucidate the relation between the initial components of waste derived fuels and the obtained reaction products.

1. INTRODUCTION

About 2.01 billion metric tons of municipal solid waste (MSW) are produced annually worldwide. By 2050, this quantity is expected to increase by up to 70% (Kaza et al., 2018). As of today, 66% of the produced waste is still dumped or landfilled, causing environmental and health risks. Current management policies try to reduce this share by focusing on prevention, reduction, and mechanical recycling. Other solid waste fractions that cannot be reused or recycled are increasingly used to produce waste derived fuels, offering an interesting alternative to conventional disposal methods. Refuse Derived Fuel (RDF) and Solid Recovered Fuel (SRF) are both produced from industrial and household wastes, the latter following specific criteria indicated in CEN/TS 15359 (Solid Recovered Fuels, 2006). Both fuels include combustible materials with high calorific values like plastics, textiles, wood, and elastomers (Garcés et al., 2016).

Among the available waste to energy technologies, gasification has received an increasing interest in the last years (Saghir et al., 2018). Gasification is a thermochemical process where carbonaceous feedstocks react above 700°C, in presence of an oxidizing agent that is fed below the stoichiometric amount needed for total oxidation. The produced syngas can be directly burnt to produce energy and heat or, after cleaning, used in a synthesis process to produce biofuels or chemicals. This process offers some

environmental advantages compared to landfilling or incineration. However, upscaling still presents some technical barriers (Centi & Perathoner, 2020).

Pyrolysis is of great interest since it is considered here as the first stage of the gasification process. During this step, the feedstock decomposes in absence of oxygen, producing volatiles in the form of light permanent gases (such as H₂, CO, CO₂, CH₄, and H₂O), and condensable hydrocarbons (tar). The remaining solid residue is known as char. Those devolatilization products participate with the oxidizing agent in the gasification step, where cracking and reforming reactions produce syngas. Distribution and product yields are highly influenced by temperature, heating rate and residence time, features that are directly related to reactor configuration (Pasel & Wanzl, 2003). Slow heating rates induce high residence times and favor char formation, while high heating rates and high temperatures favor gas products (Efika et al., 2018).

Even in the same process conditions, final products are also affected by the raw composition of the feedstock (Esmaeili et al., 2020). In previous pyrolysis/gasification studies, researchers have attempted to evaluate the influence of the feedstock using real samples of waste fuels (Efika et al., 2015; Hwang et al., 2014). However, the heterogeneous composition of waste at many levels (source, season, components, etc.), as well as the different operations to which it is subjected during the production of derived fuels, makes it really challenging to apply the results to other fuels that

* Corresponding author:
Oscar Sosa Sabogal
email: oscar.sosasabogal@cea.fr

may contain different feed compositions.

In this regard, several authors have conducted studies of the individual typical components found in waste (Chhabra et al., 2020; Meng et al., 2015; Zhou, Long, et al., 2015), most of them by thermogravimetric analysis (TGA). Mass loss curves and kinetic parameters are well known, yet few studies report the performance and composition of pyrolytic products. Moreover, the reaction conditions found in TGA and other laboratory-scale devices used in these works are not representative of full-scale gasifiers, where high heating rates and good gas-solid contact are achieved (Cortazar et al., 2020). Induction heating is an interesting alternative without the shortcomings of other conventional methods. Shorter heating times, better control, and better temperature distribution could be achieved in comparison with electrical resistance based furnaces as seen in the work of (Mishra et al., 2019). Nevertheless, the literature regarding pyrolysis or gasification applications at laboratory scale using this method is very scarce.

This work focuses on the distribution and characterization of products of pyrolysis, representing the first decomposition step of gasification of SRF in a lab scale reactor, especially developed to emulate fast pyrolysis conditions. The conversion of "model materials", selected to represent the main fractions found in solid waste, is of particular interest. Results will contribute to the development and validation of accurate prediction models, providing a better insight on the influence of feedstock on the reaction products.

1.1 Abbreviations

- MSW: Municipal Solid Waste
- RDF: Refuse Derived Fuel
- SRF: Solid Recovered Fuel
- TGA: Thermogravimetric Analysis
- LDPE: Low-Density Polyethylene
- PET: Polyethylene Terephthalate
- FID: Flame Ionization Detector
- NDIR: Non-Dispersive Infrared Detector

2. MATERIALS AND METHODS

2.1 Feedstock

A solid recovered fuel (SRF) sample produced from industrial and household waste was used for the tests. Typical discarded products from daily life were also collected and used as model materials.

Plastic samples consisted in low-density polyethylene (LDPE, referred here as PE), while polyethylene terephthalate (PET) was obtained from plastic water bottles. Wood sample consisted in pellets made from beech sawdust, and cardboard sample was obtained from packaging boxes. All materials were shredded up to particles between 1 and 3 mm, and then characterized by ultimate and proximate analyses. Moisture content was determined at 105°C while ash content was measured at 815°C according to the ISO 1171 standard. Elemental composition (C, H, N, S) was determined with an ELEMENTAR analyzer. Composition and characteristics for each sample are listed in Table 1.

2.2 Experimental setup and procedure

A laboratory scale unit was specifically designed to conduct pyrolysis and gasification tests under some reaction conditions that are representative of fast pyrolysis in a fluidized bed reactor. These characteristics include high heating rates, a temperature range between 700 and 900°C, and good gas/solid contact. Gas residence times inside the hot zone of the reactor are usually between 0.5 and 10 s for bubbling fluidized bed reactors, and between 0.5 and 1 s for circulating fluidized bed reactors (Marshall et al., 2014). For the present device, a residence time between 1 and 5 s was targeted, long enough to enable significant volatile secondary reactions. The experimental setup is shown in Figure 1.

A stainless-steel tube (560 mm in height, 30.15 mm in internal diameter) was externally heated by induction, with a water-cooled copper coil inductor (420 mm long) connected to a 12kW electrical generator (HFP 12, EFD induction GmbH). Setpoint temperature was adjusted with a PID

TABLE 1: Characteristics of SRF and model materials.

[wt%]	SRF	PE	PET	Wood	Cardboard
Moisture (ar.)	5.17%	0.29%	0.5%	7.30%	6.69%
Ash (db.)	16.38%	0%	0%	2.53%	8.80%
C (db.)	48.00%	85.50%	63.00%	46.80%	43.60%
H (db.)	6.00%	13.90%	5.50%	6.70%	6.00%
O (db by diff.)	26.68%	0.50%	31.50%	43.77%	41.10%
N (db.)	1.33%	0.02%	0.02%	0.02%	0.03%
S (db.)	0.47%	0.03%	0.02%	0.1%	0.13%
Cl (db.)	1.14%	-	-	-	-

ar: as received, db: dry basis

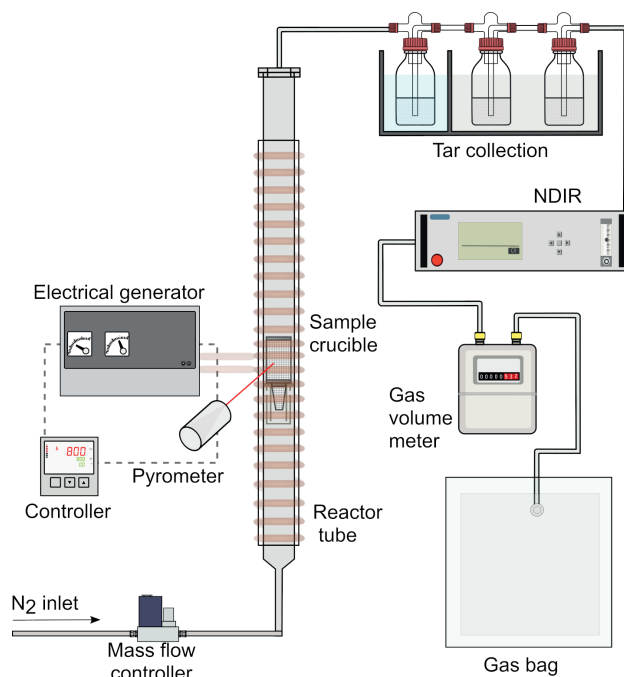


FIGURE 1: Schematic diagram of the lab scale device.

controller (GEFRAN 2500), connected to a two-color optical pyrometer (Impac IGAR 6). Its spot is pointing on the tube surface, at the height of the sample crucible.

The crucible was filled with 2 to 3g of dried feedstock (24h at 105°C) and placed in the isothermal zone of the reactor. N₂ (0.5 L/min) was used to flush the entire system and to carry the produced volatiles. The reactor was heated up to 800°C, with a heating rate of about 70°C/s, and then held at this temperature for about 20 minutes. Previous tests showed that the temperature was homogeneous in the sample, with a maximum temperature difference of 10°C.

Produced gases flowed upwards in the reactor top section and the outlet line, both traced and insulated to keep temperature above 250°C, and thus to prevent tar and water condensation before reaching the tar traps. Five gas washing bottles filled with 2-propanol were used to collect gas condensable species (water and tar) present in the stream. Glass wool and glass beads were used to improve the contact area between the gas and the solvent. The first two bottles were installed in an ice bath at 0°C, while the other three were immersed in a carbonic ice and 2-propanol bath at -70°C. Collected condensable hydrocarbons were sampled and subsequently analyzed by using a gas chromatograph system. Species were identified by the means of mass spectrometry coupled with a flame ionization detector for the quantification (GC-FID, Agilent 7890A).

Non-condensable gases were analyzed online using a Non-Dispersive Infrared Detector (NDIR Model 30, CAI), able to record CO, CO₂ and CH₄ concentrations every 1 s. The volume of the outlet gas was measured by a gas meter placed at the outlet of the NDIR analyzer. Permanent gases were then collected in a Tedlar bag to be further analyzed by micro gas chromatography (Agilent 3000A). Remaining solid products in the sample crucible were weighed and collected once the reactor had cooled down to ambient temperature.

3. RESULTS AND DISCUSSION

3.1 Carbon distribution

The carbon distribution among the pyrolysis products, namely gas, tar and char, is shown in Figure 2. These yields are calculated as the ratio of carbon in each product per mass of initial carbon in the dry ash free feedstock. Considered gas species consisted in H₂, CO, CO₂, CH₄, C₂ hydrocarbons (C₂H₄, C₂H₆, C₂H₂), C₃ hydrocarbons (C₃H₆, C₃H₈), benzene (C₆H₆) and toluene (C₇H₈). The tar fraction is defined here as all condensable organic compounds with a higher molecular weight than toluene.

For every feedstock, carbon was mostly converted into gas, as the high temperature and heating rate enhanced the thermal degradation of primary volatiles (E. C. Efiika et al., 2015). The carbon fraction in solid was under 2 wt% for PE, and corresponds to carbon deposits in the walls of the sample crucible. Cellulosic materials (wood and cardboard) showed high C content in the solid fraction. Char yield was greater in the former case, since wood has a higher lignin content than cardboard, which is composed mostly of cellulose. Lignin presents a more complex molecular structure and high fixed carbon content, which makes it more thermally stable (Muley et al., 2016). The highest fractions of carbon in tar were found for SRF (12 wt%), followed by PE (10.7 wt%) and PET (7.4 wt%). Plastic content in waste fuels improves gas yield, but also increases tar content in syngas as observed here and also by other authors (Wilk & Hofbauer, 2013; Zaccariello & Mastellone, 2015). C-balance closure ranged between 75 and 90%. Deviations could be attributed to the volatilization of light compounds from the tar traps, and undetected species that were not quantified in the FID analysis.

3.2 Solid residue (char)

The chemical composition of the solid residue obtained after pyrolysis tests is listed in Table 2. The oxygen

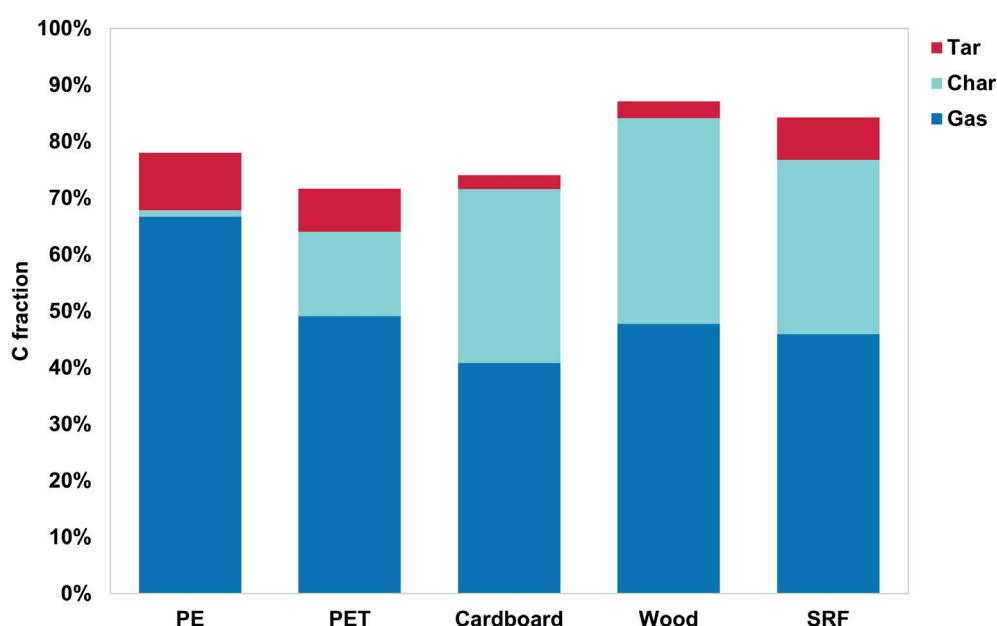


FIGURE 2: Carbon distribution from SRF and its components to pyrolysis products at 800°C.

content was calculated by difference, and the ash content was determined also at 815°C. No char was observed in the case of PE. The solid residue from SRF and cardboard pyrolysis showed the lowest amount of carbon, and high ash contents, which are related to the inorganics and inert materials from the original feedstock.

3.3 Gas composition

Yield and volumetric composition of the produced gases from SRF and model materials pyrolysis, are displayed in Table 3. Low density polyethylene showed the highest yield of gas. Its long-branched structure, follows a random chain scission mechanism during its devolatilization, giving high yields of hydrogen, light olefins (mostly C2 hydrocarbons) and no char as stated by other authors (Al-Salem et al., 2017; Block et al., 2019). Gas yields values were similar for SRF and the other studied materials.

CO and CO₂ were the main produced gas species in PET pyrolysis due to presence of oxygen groups from its monomer. Hydrocarbon production was much lower compared to PE, as observed also in the work of (Honus et al., 2018). Biogenic feedstock (wood, cardboard), which have an important O content, yielded the highest concentrations of oxygenated gases. CO and CO₂ contents were higher for cardboard case, while methane content was higher for wood. Raw materials used in the production of paper and cardboard are submitted to delignification processes, so their cellulose content can go up to 99% (Runchal et al., 2018). (Yang et al., 2007), suggested that cracking of carbonyl functional groups of cellulose gave high CO yields, while the degradation of aromatic rings and methoxyl groups in lignin enhanced CH₄ yields.

(Win et al., 2020) conducted flash pyrolysis experiments of wood pellets and polyethylene in an electric furnace at 900°C. In their results, yield of H₂ was higher for wood pellet (25% vol) than that for polyethylene (20% vol), which differs from the findings presented here. They attributed those results to water-gas reactions, enhanced from the moisture content of wood pellets. In our study, samples were dried prior to the experiments, and pyrolysis temperature was lower, hence these reactions were not favored. Yields of methane and ethylene were close to our results for both materials.

Regarding the composition of the gas produced in the pyrolysis of the SRF, CO was the major gas component (30%), followed by CH₄ (23%) and H₂ (22%). Gas distribution after SRF pyrolysis was intermediate between different materials, being plastics responsible for the high yields of CH₄ and C₂ hydrocarbons while biogenic materials for the

majority of oxygenated compounds. In general, the results found in this study followed the same trends observed in RDF pyrolysis tests conducted by other researchers at 800°C (Blanco et al., 2012; Daouk et al., 2018). It is known that the sample used here presented small fractions of other fossil derived materials such as PS and rubber, known for generate high amounts of methane during their pyrolysis (Zaini et al., 2019).

3.4 Tar Composition and content

The condensed tars were sampled and then analyzed by GC – FID. Quantified tar components were categorized according to the classification system proposed by the Energy research Center of the Netherlands (ECN) (Devi et al., 2005). This classification is based on the solubility and condensation properties of the different tar compounds, which define the downstream treatment conditions. Tars are grouped in five classes (Table 4) depending on the number of aromatic rings.

As seen in Figure 3, tar content was much higher for plastics compared to lignocellulosic materials. When comparing the relative distribution of tar products by group, wood, cardboard and PET showed high amounts of heterocyclics (class 2), due to the presence of oxygenated compounds like phenol. Class 4 species content was higher for wood than for cardboard thanks to the presence of lignin, which is known to yield higher amounts of naphthalene (Zhou, Wu, et al., 2015). The primary products of PE pyrolysis consist mostly in alkenes, which react via Diels-Alder reactions producing single ring tar compounds like styrene and indene, and some 2- ring tars like naphthalene. PET monomer contains aromatic rings in its structure, which influence the formation of small amounts of heavy polyaromatic species (3 or more rings). SRF showed the highest tar yield, with high contents of heavy aromatics. This may be attributed to the presence of monomers with aromatic rings as evidenced with PET, but also others such as polystyrene.

4. CONCLUSIONS

Pyrolysis tests of SRF and selected modeled materials were carried out in a specifically developed induction heated reactor at 800°C. The most common polymers in waste (PE and PET) were the model materials for the plastic fraction, while beech wood and cardboard were

TABLE 2: Chemical composition of the remaining solid residue after pyrolysis at 800°C.

wt %	%C	%H	%N	%S	%O	%Ash
SRF	35.97	1.32	0.96	1.05	11.60	49.1
PET	85.17	1.49	0.22	0.14	12.98	0.98
Wood	77.8	1.01	0.64	0.12	12.56	7.87
Cardboard	45.13	0.73	0.37	0.14	14.3	39.33
PE	No char					

TABLE 3: Gas composition from pyrolysis at 800°C of SRF and model materials.

% vol	SRF	PE	PET	Wood	Cardboard
H2	22.1%	29.5%	24.8%	19.8%	20.5%
CO	30.1%	0.6%	31.6%	44.2%	48.7%
CO2	9.8%	0.4%	22.9%	12.0%	12.7%
CH4	22.4%	25.6%	8.6%	18.2%	13.2%
C2	9.5%	33.0%	1.8%	4.5%	4.2%
C3	0.1%	1.4%	0.0%	0.0%	0.1%
BTX	5.1%	4.0%	10.3%	0.9%	0.3%
Gas yield[L/g daf]	0.4215	0.7815	0.4463	0.4439	0.4203

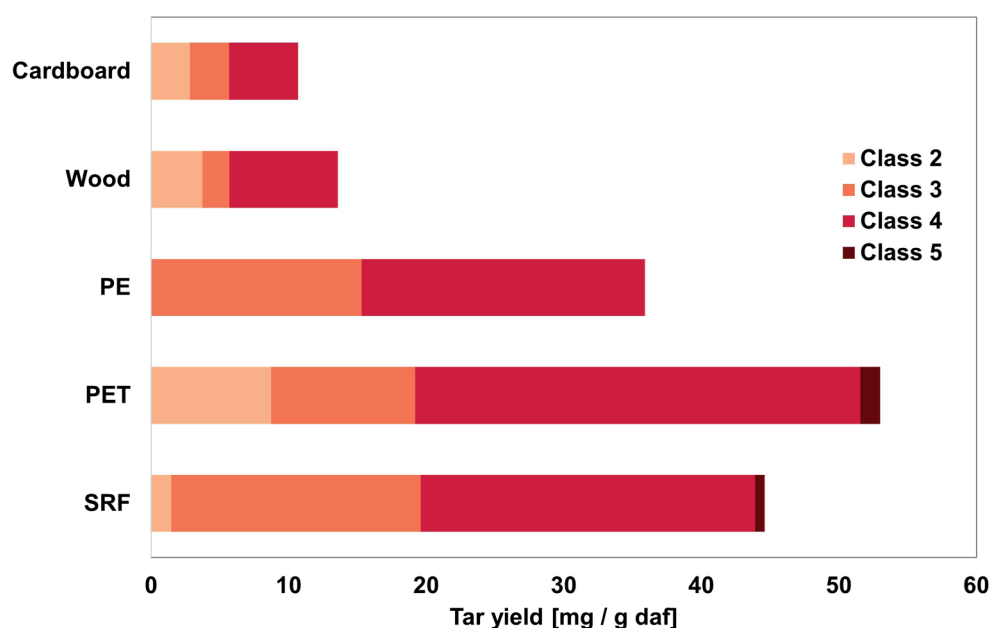


FIGURE 3: Composition of produced tar from the pyrolysis of SRF and model materials at 800°C.

representative of lignocellulosic biomass. The reaction products were quantified and analyzed, and the composition of the produced gas was reported for each material. For all cases, carbon conversion to gas was higher than 40%, as the high heating rates favored secondary reactions like the cracking of volatile products to lighter stable gases. Even if the elemental composition of some materials was very similar, differences were found in the gaseous and condensable products, showing that macromolecular structure can influence the final product distribution. These results will be helpful for the development of predictive models.

However, to have a better insight of the thermochemical conversion of solid waste and its derived fuels, additional studies must be conducted. Co-pyrolysis of waste is of particular interest, as synergetic effects may appear between the different materials modifying the final distribution of the products. Additional tests in partial oxidation conditions are also important to explore the impact of the gas phase reactions in the syngas and tar composition.

REFERENCES

- Al-Salem, S. M., Antelava, A., Constantinou, A., Manos, G., & Dutta, A. (2017). A review on thermal and catalytic pyrolysis of plastic solid waste (PSW). *Journal of Environmental Management*, 197, 177–198. <https://doi.org/10.1016/j.jenvman.2017.03.084>
- Blanco, P. H., Wu, C., Onwudili, J. A., & Williams, P. T. (2012). Characterization of Tar from the Pyrolysis/Gasification of Refuse Derived Fuel: Influence of Process Parameters and Catalysis. *Energy & Fuels*, 26(4), 2107–2115. <https://doi.org/10.1021/ef300031j>
- Block, C., Ephraïm, A., Weiss-Hortala, E., Minh, D. P., Nzihou, A., & Vandecasteele, C. (2019). Co-pyrolysis of Plastics and Biomass, a Review. *Waste and Biomass Valorization*, 10(3), 483–509. <https://doi.org/10.1007/s12649-018-0219-8>
- Centi, G., & Perathoner, S. (2020). Chemistry and energy beyond fossil fuels. A perspective view on the role of syngas from waste sources. *Catalysis Today*, 342, 4–12. <https://doi.org/10.1016/j.cattod.2019.04.003>
- Chhabra, V., Bamberg, K., Bhattacharya, S., & Shastri, Y. (2020). Thermal and in situ infrared analysis to characterise the slow pyrolysis of mixed municipal solid waste (MSW) and its components. *Renewable Energy*, 148, 388–401. <https://doi.org/10.1016/j.renene.2019.10.045>
- Cortazar, M., Lopez, G., Alvarez, J., Arregi, A., Amutio, M., Bilbao, J., & Olazar, M. (2020). Experimental study and modeling of biomass char gasification kinetics in a novel thermogravimetric flow reactor. *Chemical Engineering Journal*, 396, 125200. <https://doi.org/10.1016/j.cej.2020.125200>

TABLE 4: Tar classification according to ECN.

Tar class	Class Name	Properties	Compounds considered in this study
1	GC-undetectable	Very heavy tars, cannot be detected by GC	
2	Heterocyclic	Tars containing heteroatoms; highly water soluble compounds	Phenol, Benzofuran
3	Light aromatics	Usually light hydrocarbons with one ring; do not pose a problem regarding condensability and solubility	Ethylbenzene, Styrene, <i>p</i> -Methylstyrene,
4	Light polyaromatics	Two and three ring compounds; condense at low temperature even at very low concentration	Indene, Naphthalene, methyl-naphthalene, Biphenyl, Acenaphthylene, Acenaphthene, Dibenzofuran, Fluorene
5	Heavy polyaromatics	Larger than three-rings; these components condense at high temperatures at low concentrations	Fluoranthene, pyrene

- Daouk, E., Sani, R., Pham Minh, D., & Nzihou, A. (2018). Thermo-conversion of Solid Recovered Fuels under inert and oxidative atmospheres: Gas composition and chlorine distribution. *Fuel*, 225, 54–61. <https://doi.org/10.1016/j.fuel.2018.03.136>
- Devi, L., Ptasiński, K. J., Janssen, F. J. J. G., van Paasen, S. V. B., Bergman, P. C. A., & Kiel, J. H. A. (2005). Catalytic decomposition of biomass tars: Use of dolomite and untreated olivine. *Renewable Energy*, 30(4), 565–587. <https://doi.org/10.1016/j.renene.2004.07.014>
- Efika, C. E., Onwudili, J. A., & Williams, P. T. (2018). Influence of heating rates on the products of high-temperature pyrolysis of waste wood pellets and biomass model compounds. *Waste Management*, 76, 497–506. <https://doi.org/10.1016/j.wasman.2018.03.021>
- Efika, E. C., Onwudili, J. A., & Williams, P. T. (2015). Products from the high temperature pyrolysis of RDF at slow and rapid heating rates. *Journal of Analytical and Applied Pyrolysis*, 112, 14–22. <https://doi.org/10.1016/j.jaap.2015.01.004>
- Esmaeili, V., Ajalli, J., Faramarzi, A., Abdi, M., & Gholizadeh, M. (2020). Gasification of wastes: The impact of the feedstock type and co-gasification on the formation of volatiles and char. *International Journal of Energy Research*, n/a(n/a), Article n/a. <https://doi.org/10.1002/er.5121>
- Garcés, D., Díaz, E., Sastre, H., Ordóñez, S., & González-LaFuente, J. M. (2016). Evaluation of the potential of different high calorific waste fractions for the preparation of solid recovered fuels. *Waste Management*, 47, 164–173. <https://doi.org/10.1016/j.wasman.2015.08.029>
- Honus, S., Kumagai, S., Fedorko, G., Molnár, V., & Yoshioka, T. (2018). Pyrolysis gases produced from individual and mixed PE, PP, PS, PVC, and PET—Part I: Production and physical properties. *Fuel*, 221, 346–360. <https://doi.org/10.1016/j.fuel.2018.02.074>
- Hwang, I.-H., Kobayashi, J., & Kawamoto, K. (2014). Characterization of products obtained from pyrolysis and steam gasification of wood waste, RDF, and RPF. *Waste Management*, 34(2), 402–410. <https://doi.org/10.1016/j.wasman.2013.10.009>
- Kaza, S., Yao, L., Bhada-Tata, P., & Van Woerden, F. (2018). What a Waste 2.0: A Global Snapshot of Solid Waste Management to 2050. The World Bank. <https://doi.org/10.1596/978-1-4648-1329-0>
- Marshall, A. J., Wu, P. F., Mun, S.-H., & Lalonde, C. (2014). Commercial application of pyrolysis technology in agriculture. *American Society of Agricultural and Biological Engineers Annual International Meeting 2014, ASABE 2014*, 5, 3868–3886.
- Meng, A., Chen, S., Long, Y., Zhou, H., Zhang, Y., & Li, Q. (2015). Pyrolysis and gasification of typical components in wastes with macro-TGA. *Waste Management*, 46, 247–256. <https://doi.org/10.1016/j.wasman.2015.08.025>
- Mishra, H., Patidar, B., Pante, A. S., & Sharma, A. (2019). Mathematical modelling, simulation and experimental validation of resistance heating and induction heating techniques for E-waste treatment. *IET Electric Power Applications*, 13(4), 487–493. <https://doi.org/10.1049/iet-epa.2018.5535>
- Muley, P. D., Henkel, C., Abdollahi, K. K., Marculescu, C., & Boldor, D. (2016). A critical comparison of pyrolysis of cellulose, lignin, and pine sawdust using an induction heating reactor. *Energy Conversion and Management*, 117, 273–280. <https://doi.org/10.1016/j.enconman.2016.03.041>
- Pasel, C., & Wanzl, W. (2003). Experimental investigations on reactor scale-up and optimisation of product quality in pyrolysis of shredder waste. *Fuel Processing Technology*, 80(1), 47–67. [https://doi.org/10.1016/S0378-3820\(02\)00187-X](https://doi.org/10.1016/S0378-3820(02)00187-X)
- Runchal, A. K., Gupta, A. K., Kushari, A., De, A., & Aggarwal, S. K. (2018). *Energy for Propulsion A Sustainable Technologies Approach*. Springer Singapore: Imprint: Springer. <https://link.springer.com/10.1007/978-981-10-7473-8>
- Saghir, M., Rehan, M., & Nizami, A.-S. (2018). Recent Trends in Gasification Based Waste-to-Energy. In Y. Yun (Ed.), *Gasification for Low-grade Feedstock*. InTech. <https://doi.org/10.5772/intechopen.74487>
- Solid Recovered Fuels: Specification and Classes: CEN/TS 15359 Technical Specification: English Version. (2006). European Committee for Standardization.
- Wilk, V., & Hofbauer, H. (2013). Conversion of mixed plastic wastes in a dual fluidized bed steam gasifier. *Fuel*, 107, 787–799. <https://doi.org/10.1016/j.fuel.2013.01.068>
- Win, M. M., Asari, M., Hayakawa, R., Hosoda, H., Yano, J., & Sakai, S.-I. (2020). Gas and tar generation behavior during flash pyrolysis of wood pellet and plastic. *Journal of Material Cycles and Waste Management*, 22(2), 547–555. <https://doi.org/10.1007/s10163-019-00949-8>
- Yang, H., Yan, R., Chen, H., Lee, D. H., & Zheng, C. (2007). Characteristics of hemicellulose, cellulose and lignin pyrolysis. *Fuel*, 86(12), 1781–1788. <https://doi.org/10.1016/j.fuel.2006.12.013>
- Zaccariello, L., & Mastellone, M. L. (2015). Fluidized-Bed Gasification of Plastic Waste, Wood, and Their Blends with Coal. *Energies*, 8(8), 8052–8068. <https://doi.org/10.3390/en8088052>
- Zaini, I. N., García López, C., Pretz, T., Yang, W., & Jönsson, P. G. (2019). Characterization of pyrolysis products of high-ash excavated-waste and its char gasification reactivity and kinetics under a steam atmosphere. *Waste Management*, 97, 149–163. <https://doi.org/10.1016/j.wasman.2019.08.001>
- Zhou, H., Long, Y., Meng, A., Li, Q., & Zhang, Y. (2015). Thermogravimetric characteristics of typical municipal solid waste fractions during co-pyrolysis. *Waste Management*, 38, 194–200. <https://doi.org/10.1016/j.wasman.2014.09.027>
- Zhou, H., Wu, C., Onwudili, J. A., Meng, A., Zhang, Y., & Williams, P. T. (2015). Polycyclic aromatic hydrocarbons (PAH) formation from the pyrolysis of different municipal solid waste fractions. *Waste Management*, 36, 136–146. <https://doi.org/10.1016/j.wasman.2014.09.014>